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Parabolic trough concentrator design, characterization, and application: solar wastewater purification targeting textile industry dyes and pharmaceuticals—techno-economic study

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Abstract

Solar energy, along with other renewable resources, has the potential to be a major contributor to solving environmental issues in the future, as illustrated by the most recent advancements in solar photocatalytic technology. Indeed, wastewater treatment using a parabolic solar collector for industrial processes is gaining ground owing to improved system performance and economic benefits. The fabricated parabolic trough collector (PTC) incorporates reflective, parabolic panels that focus solar energy onto a transparent tube positioned along the parabolic focal line, where solar-powered photochemical reactions occur. This study investigated the design, implementation, and effectiveness of a concentrated sunlight system for removing industrial dyes and emerging large-use pharmaceutical contaminants in the presence of H_2O_2 at a small demonstrator scale (10 L/h). A spectrophotometric assessment revealed that subjecting Remazol Brilliant Blue (RBB, 60 ppm) and ciprofloxacin (CIP, 10 ppm) to irradiation in the presence of 0.1 M H_2O_2 (RBB) or 0.01 M H_2O_2 (CIP) for 3 h resulted in a degradation rate exceeding 60% and 80%, respectively. Furthermore, the total organic content (TOC) analysis indicates a very high total removal yield for RBB. On these bases, a techno-economic analysis is produced, and economic viability is discussed. The data reveal that the annual costs for water treatment, considering investment, electricity, and catalyst expenses over a 12-month period are significantly lower for our PTC-based prototype than for a comparable artificial UV-based equipment.

Keywords Parabolic trough collector \cdot Wastewater treatment \cdot Industrial dyes \cdot Antibiotics \cdot Concentrated sunlight system \cdot Techno-economic assessment

Introduction

The over-dependence on fossil fuel as a primary energy source has adverse effects on health, the environment, and climate. To address these problems, researchers are progressively finding ways towards the exploitation of energy from renewable sources. Numerous companies on a global scale manufacture solar technology and provide innovative solutions for cleaner and reduced-cost energy. The

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implementation of solar power is extensively employed across various domains including the provision of water heating and electricity generation. However, while a great deal of lab-scale research has been carried out on solar photocatalysis for water purification, prototype level studies are still relatively rare (Suresh et al. 2014; Krüger et al. 2008; Fernández-García et al. 2010; Pitz-Paal and Second 2014; Ydrissi et al. 2020; Esteban García et al. 2021).

Considering the application to wastewater purification, the use of PTCs in solar concentrating technology is prevalent, as an eco-friendly and sustainable strategy (Tanveer and Tezcanli Guyer 2013; Bhagwati et al. 2023; Saini et al. 2023). Over the long term, this system has the potential to be economically efficient. Following the establishment of the initial setup, operational expenses typically prove to be less than those associated with traditional, energy-intensive methods of wastewater treatment. Moreover, solar energy concentration systems can be structured in a modular

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manner, facilitating scalability and adaptability to various industrial environments (Saini et al. 2023; Fredriksson et al. 2021).

The substantial load of pollutants in industrial wastewater, along with its possible repercussions on water quality and public health, designates it as a significant environmental issue (Singh et al. 2023). Antibiotics stand out as emerging contaminants (Michael et al. 2013; Hou et al. 2020; Ferro et al. 2015). This problem is made urgent by the recent pandemic situation that has led to a massive increase in the use of especially antibacterial drugs for which conventional water treatment methods have limited removal capacity. The release of these pollutants into the environment involves risks to the ecosystem and health, due to both direct exposure and the generation of resistant bacteria (Harrington et al. 2022; Wang et al. 2022; Huang et al. 2022; Al-Riyami et al. 2018). Beyond the immediate environmental and health considerations, there exist concomitant economic implications. The emergence of antibiotic-resistant bacterial strains has the potential to engender heightened healthcare expenditures, precipitated by the imperative for more intricate and specialized treatments (Saeed et al. 2023).

Also, the discharge of textile industry waste products, specifically dyes, is a significant contributor to water pollution (Islam and Mostafa 2019; Ben et al. 2021; Lellis et al. 2019; Hossain et al. 2018). Notably, vinyl sulfone dyes-a type of reactive dye commonly applied to cotton, silk, and wool known as Remazol dyes, after their original trademark—are commonly employed (Lewis 2011). Despite their effectiveness for dyeing, vinyl sulfone dyes are highly toxic and pose a significant threat to both aquatic and human life. Once released into water bodies, vinyl sulfone dyes can persist for long periods of time and cause skin irritation, damage to internal organs, as well as lasting damage to ecosystems. Some dyes are recognized as potential cancercausing agents, suggesting a risk of cancer development upon exposure. (Lellis et al. 2019). Additionally, certain dye compounds can disrupt the endocrine system, affecting hormonal balance in both humans and wildlife. This interference can result in various negative effects on reproductive and developmental processes (Kumar et al. 2008). Moreover, due to their specific chemical composition, the degradation of these dyes presents a particular challenge for conventional physical and chemical treatment methods (Gouvêa et al. 2000; Niebisch et al. 2010; Ara et al. 2013; Sanmuga Priya et al. 2015).

Advanced oxidation processes (AOPs) (Deng and Zhao 2015), particularly via photocatalysis, have been proposed as an alternative for the treatment of recalcitrant wastewater. These processes are characterized by the production of radical species, typically the 'OH radical, on semiconductor materials in contact with water, following the absorption of photons (Sirés et al. 2014; Rodríguez et al. 2004).

Alternatively, radicals can be produced by a sacrificial source such as H_2O_2 either by simple photochemical dissociation or in photo-Fenton reactions (Lewis 2011). In particular, the light-induced homolytic dissociation of H_2O_2 can provide an efficient way to generate hydroxyl radicals (Liu et al. 2017) in water. These radicals then proceed to react with most organic pollutants resulting, in principle, into their complete mineralization (Venkatadri and Peters 1993). This procedure was adopted in our study to assess the effectiveness of a concentrated sunlight system in removing harmful industrial contaminants.

Two general forms of concentrating solar collectors exist, namely point-focus systems-encompassing parabolic dish collectors and central receiver towers-and line-focus systems, including linear Fresnel and PTC (Ydrissi et al. 2020; Xie et al. 2011; Coelho et al. 2014; Golli et al. 2021). Parabolic trough systems are the most deployed and advanced ones (Ydrissi et al. 2020; Alamr and Gomaa 2022; Cavallaro 2009). The fundamental concept underlying this technology entails the focusing of solar radiation onto a tubular receptor in which, wastewater may circulate so that the solar radiation energy will be an input for photochemical reactions. Indeed, the elementary parts of a PTC are the parabolic trough shape, which has a symmetrical section of a parabola around its vertex, and a reflective surface that concentrates the radiation on the tubular receiver located in the focus line of the parabola (Braham and Harris 2009).

This article outlines a straightforward design and its implementation for producing a solar concentrator with the aim of assessing the technical performance and economic viability of a PTC for solar wastewater treatment of two case-study compounds: a large-use antibiotic (Ciprofloxacin) and a vinyl sulfone industrial dye (Remazol Brilliant Blue). The first part describes the design of the adopted parabola configuration together with a detailed description of the main components. The second part presents the characterization of the mirror by means of reflectivity measurements and the results of tests including the evaluation of the manufactured PTC efficiency for the degradation of the two different types of pollutants (dyes, and pharmaceuticals) using H₂O₂ as a sacrificial hydroxyl radicals source. Finally, based on these data, a techno-economic assessment discussing the viability of this technology is presented.

Experimental and methodology

Reactor

Given that this work concerns the use of solar concentration for the purification of contaminated water, and that the size of the final apparatus must be on the scale of a small industrial prototype, we start with the description of the receiver hosting the treatment processes. Once the geometry of this component has been defined, the technical aspects of the concentrator will be presented, which must be suitable for the type of reactor.

The amount of circulating water is established to be 5 L moving from a tank to the solar focus, where a quartz tube (Helios quartz, quality class NHI[®]-1100) is positioned with a volumetric capacity of 2.85 L and acting as a reactor. The quartz cylinder is closed at both ends by two Teflon lids, with the one at the bottom containing an inlet port for liquid circulation in polyamide pipes (external diameter 8 mm, internal diameter 6 mm). The top lid is upheld by a T-junction, with one of the ports

housing a thermocouple for temperature monitoring, and another port being the liquid outlet. Solution circulation is achieved using a dosing peristaltic pump (Seko, model KRFF0210) sustaining a flow rate of 10 L/h (Fig. 1).

The reactor is fixed on the mechanical structure of the solar concentrator. The structure is made of a ground-supported galvanized steel column, with a flat metal frame installed on top, where the parabolic mirror is mounted. The frame is moved by a two-axis automatic tracker motor system. The motors are guided by the signal of two sensors that constantly measure the maximum average irradiance and allow the motors to adjust to the best illumination throughout the day.



Fig. 1 The solar collector at the University of Trento assembled on field: the pump (A), mirror (B), receiver (C), solution reservoir (D), solar tracker (E), controller joystick (F), and circulation pipes (G)

Solar parabolic trough concentrator design

Geometric parameters of PTC

A preliminary investigation of geometry is necessary to attain optimal effectiveness (Fig. 2) and capture the maximal radiative intensity obtained at the collector aperture (Yılmaz and Mwesigye 2018; Izweik et al. 2016; Waghmare and Gulhane 2017).

The collector's profile is established using the spatial coordinates (x and y) as follows:

$$x^2 = 4yf. \tag{1}$$

In Eq. (1), f is the focal length, which defines the position of the solar receiver part relative to the parabola vertex and may be calculated as:

$$f = x_{\rm LP} / 4\tan\frac{\varphi_{\rm r}}{2} \tag{2}$$

 $\varphi_{\rm r}$ denotes the rim angle which is obtained as follows:

$$\varphi_{\rm r} = \sin^{-1}(x_{\rm LP}/2Rr) \tag{3}$$



Fig. 2 Cross-section of the PTC

Rr is the radius of the rim. The local mirror radius may be calculated at any point along the collector

$$Rr = 2f/(1 + \cos\varphi_r). \tag{4}$$

The geometric characteristics of the designed collector are depicted in Table 1.

Geometric concentration ratio (GCR)

The geometric concentration ratio (GCR) for a parabolic trough collector refers to the ratio between the area that intercepts the solar radiation at the collector aperture (A_a) and the area that receives the sunlight (A_{rec}) (Rodríguez et al. 2004; Spasiano et al. 2015). A high GCR indicates that a large proportion of incident light is focused onto a smaller area, providing a more intense energy flux (Fig. 3).

The geometrical concentration ratio is given by Izweik et al. (2016); Esmaeili et al. 2023):

 Table 1
 Parameters of PTC (He et al. 2011)

| | Parameter | Value (m) |
|----------------------------|-----------------|-----------|
| Parabolic trough reflector | x _{LP} | 1.1 |
| | $y_{\rm LP}$ | 0.95 |
| | f | 0.3 |
| Receiver | D_{ai} | 0.055 |
| | Ε | 0.005 |

 x_{LP} : Length of parabola in x direction (aperture width)

 y_{LP} : Length of parabola in y direction

f: Focal length

 D_{ai} : Inner diameter of absorber tube

 D_{ao} : Outer diameter of absorber tube

E: Thickness of the absorber tube



Fig.3 A scheme representing the parabolic trough solar collector (Esmaeili et al. 2023)

$$GCR = A_a / A_{rec} = La / L\pi D_{ao}$$
⁽⁵⁾

a: aperture width, D_{ao} : outer diameter of absorber tube.

The GCR has a minimum value of 6 when considering an outer diameter of 60 mm of the receiver tube and a maximum value of 118 while taking into account the focal spot size in the center of the quartz tube, which is about 3 mm.

Manufacture and assembly of different components

Support structure Table S1 shows the schematic of subparts of PTC assembly with their respective dimensions. In addition to the base panel, the collector assembly comprises a pair of "supporting arms" to uphold the quartz tube and mirror. The "lateral apertures" are made of 10 mm thick solid aluminum and the reflectors were securely attached to the supporting arms utilizing aluminum C-profile inserts.

Parabolic reflectors Finding the ideal reflective surface for solar photochemical applications is a challenge, as it requires a balance of UV reflectivity, weather resistance, and cost-effectiveness. A good compromise on these essential criteria is electropolished anodized aluminum (Spasiano et al. 2015; Jorgensen and Govindarajan 1991; Malato et al. 2002), such as the 0.5-mm-thick anodized aluminum MIRO-SUN[®] sun provided by Alanod, Germany. This mirror features a central core in aluminum with anodized layers on both sides, an epoxy layer is deposited in the back for protective installation and a protective weatherproof coating is present on the top.

The direct and reflected spectra of the sun were acquired on the PTC solar system under ideal atmospheric conditions, as shown by the spectra in Fig. 4a. These spectra were obtained using a mini spectrophotometer TM–UV–VIS (Hamamatsu Photonics K.K. model C10082CA, with minispec EvaluationSoftware), orienting it towards the direct sun rays and toward the reflective mirror. Remarkably, the mirror largely reproduces the solar spectrum without distortions. The mirror reflectance was measured by using a double-beam spectrophotometer (VARIAN Cary 5000), and corresponding reflectance spectra as a function of wavelength is presented in Fig. 4b. The mirror's optical properties feature a reflectivity above 80% in the measured range between 330 and 800 nm.

Receiver tube The reactor tube is an enclosed medium that hinders the evaporation of volatile substances. Since the photochemical reactor envelops the operative medium, it must be transparent to UV–visible radiation and it must also withstand pressure build-ups for industrial perspectives. Therefore, the tubular-shaped reactor configuration is the most appropriate for handling and pumping water. This configuration is deemed ideal for the photoreactor setup to maintain the required pressure and flow rates needed for circulation systems.

The options for optically transmissive and UV-resistant materials are restricted. In addition, it is essential for the reactor component to be unaffected by the hazardous substances, predominantly OH radicals that may course through it. Consequently, the use of quartz for the tube (60.00 ext. dia. $\times 2.50$ thick $\times 1200$. 00 length mm) is preferred since it



Fig. 4 a Solar spectra measured on direct sunlight and reflected radiation by the mirror, b reflectance as a function of wavelength of the mirror used in this work

boasts remarkable UV transmittance and chemical resilience at high temperatures (Rodríguez et al. 2004; Qu et al. 2007; Olia et al. 2019).

Solar tracking system To improve the effectiveness of direct solar radiation collection, the platform is equipped with a solar tracking system that keeps the aperture plane perpendicular to the incident sunlight (Spasiano et al. 2015), and its concentration on the focal line (Fig. 5). The collector's dual-axial tracking system is maneuvered through both azimuthal (pertaining to an axis aligned with the north–south direction) and elevational (pertaining to an axis aligned with the east–west direction) axes. Additionally, the tracking system can be programmed to cope with adverse climatic conditions or precarious circumstances.

The tracking technology employed here is supplied by the DEGER company (DEGERtraker 3000NT) and it is originally designed to carry photovoltaic modules. The electronic part is used as provided, while the supporting frame has been adapted in-house to support the PTC instead of solar panels.

Preliminary test of solar wastewater treatment application

The manufactured PTC efficiency for the degradation of two different types of pollutants (dye and pharmaceutical) has been studied. 5 L of 60 mg/L of RBB and 10 mg/L of CIP solutions have been studied. The dyes were provided by a textile plant (ERAK, Istanbul, Turkiye) and the pharmaceutical CIP was obtained from Sigma Aldrich (Standard PHR1044). All irradiation experiments were preceded by a



Fig. 5 Sketch of the parabolic trough collector (Olia et al. 2019)

30-min homogenization stage with the PTC (i.e., not facing direct sunlight on to PTC).

The temperature in the reactor was monitored by an insulated T-type thermocouple immersed in the studied liquid and kept under 70 °C by pausing the experiment if needed. This was done simply by moving the PTC out of focus and allowing the solution to cool underflow for a few minutes. This was only needed during experiments with the intensely colored RBB, while temperature never reached more than 45 °C with CIP. Solar irradiance values were measured with a pyrheliometer (Kipp&Zonnen, model CHP1) and they ranged between the values of 600 and 700 W/m². The PTC is installed in Trento, northern Italy (46.0664°N, 11.1507°E) and experiments were carried out between 10:30 and 14:00 local time, during the months of March and July 2023. Every 15 min in the first hour of the experiment, 1 ml of the studied fluid was collected for the acquisition of the absorbance spectrum by a double-beam spectrophotometer (VARIAN Cary 5000 UV-VIS-NIR), to evaluate the reaction kinetically. After that, every 30 min, 1 ml of the solution was analyzed. Note that the degradation can be simulated using a pseudo-first order reaction $\ln (Ci/C) = k \times t$, authenticated by Langmuir-Hinshelwood for the decreasing amount of pollutant. The constant rate, k, can be obtained from the graphed linear correlation of $\ln(C_0/C)$ versus t (Gopalakrishnan et al. 2011). Qualitative measurements of CO_2 evolution were done using an IR sensor (COZIR GSS 5% model) and the Gaslab software for data acquisition. To evaluate the contribution of direct photolysis, blank experiments in the absence of H_2O_2 were also performed.

Textile industry dye

RBB dye degradation was examined in the presence of 0.01 M, 0.1 M, and 1 M H_2O_2 (130 v/v) under concentrated sunlight, as reported in Fig. 6a. It was observed that hydrogen peroxide was not activated in the absence of irradiation and the concentration of the dye remained unaltered for the duration of 30 min under dark.

At 0.01 M H₂O₂ concentration, the degradation of the RBB under light achieved only about 20% in 3 h under concentrated sunlight. Increasing the concentration of H₂O₂ to 1 M increased decolorization efficiencies, which resulted in 53% degradation in 1 h. With 0.1 M H₂O₂ concentration and 3-h reaction time it reached about 61% degradation. The rate constant *k* for the tests at H₂O₂ concentrations of 1 M, 0.1 M, and 0.01 M were 0.01178, 0.00446, and 0.00114 min⁻¹, respectively (Fig. 6b). In the case of direct photolysis, in the absence of radical source, the observed discoloration was negligible.

Additionally, CO_2 emissions were measured and reported (Figure S1) to qualitatively confirm the occurrence of mineralization. Indeed, [•]OH radicals may lead



Fig. 6 a Combined effects of concentrated sunlight and $H_2O_2 0.01$ M, 0.1 M, and 1 M H_2O_2 on decolorization efficiency of RBB. **b** Corresponding kinetic constants for photodegradation of RBB under sunlight irradiation.

to the total oxidation of organic compounds with CO_2 as the final product (Fendrich et al. 2018). Total organic content analysis confirms a mineralization yield of 72% with 0.1 M H₂O₂.

Ciprofloxacin (CIP)

For the CIP antibiotic, a photolysis test with concentrated sunlight resulted in 34% degradation, while with just 0.01 M of H_2O_2 , the degradation resulted in 88% (Fig. 7a). Moreover, when using only non-concentrated (non-CC) light, without focusing, the degradation of CIP was negligible



Fig. 7 a Tests on PTC for degradation of the pharmaceutical: CIP: (red). Photolysis by only concentrated sunlight, (blue) concentrated sunlight $+0.01M H_2O_2$, (black) non-concentrated sunlight, **b** Corresponding kinetic constant for photodegradation of CIP.

compared to the results obtained when exposing the solution to concentrated (CC) sunlight. This last test highlights the role of the parabolic solar concentrator. The *k* for CIP was 0.00965 min⁻¹ (Fig. 7b). In this case, TOC analysis shows only an 8% mineralization yield. It is worth highlighting here that the aim of antibiotic residuals treatment is generally inactivation rather than mineralization, since the ultimate goal is that of preventing the arising of resistance in bacteria. To this end, the Ciprofloxacin structural change demonstrated by the disappearance of the UV–visible absorption might be sufficient. Further in-depth studies on the residual antibiotic activity of the treated water are underway to verify it.

Techno-economic assessment

An economic assessment of the small-scale prototype is presented in this section to elucidate the water treatment costs involved in the operation of the proposed system. This assessment takes into account a comparison between (1) our PTC-based process versus an artificial UV-light apparatus with 1000 W light power, which we consider with equivalent or higher efficiency as the proposed PTC, both in continuous circulation flow rate of 10 L/h and (2) a process using 0.1 M H₂O₂ as sacrificial agent versus the benchmark colloidal photocatalyst Evonik P25 TiO₂ nanoparticles employed with 1 g/L concentration. Specifically, the study provides estimates in such conditions for 3-h treatments achieving 80% removal for RBB-based on the investment cost, electricity usage, and consumable materials pricing (Giménez et al. 2019).

Table 2 lists the components for both setups and the experimental conditions. A 12-month period of water treatment was chosen as the time reference with 3 weekly water treatments, meaning nearly 150 runs in a year, each of them with 5 L of 60 ppm concentration of RBB dye present in the batch.

Table 3 details the investment costs for both the PTC and the UV apparatus: the parts acquisition price, and the manufacturing and machined items costs. The 2-axis automatic solar tracking and the UV lighting prices are based on actual market quotations (September/2023). Calculation results show yearly costs of \notin 1230 for the PTC while \notin 3960 for the UV source. Shorter lifetimes (maximum 2 years) of UV lamps influence significantly (88%) the higher investment for the UV system requiring periodical changes, while the PTC tracker's longer-term life motor provides an advantage.

Table 4 details the energy costs for the solar PTC of this work and UV infrastructure for 1 year of water treatments. For this case, it was considered a single unit of treatment (3 h) related to the power consumption of each electrical part from both systems (KWh). The average electricity price of €0.4/KWh was attributed to considering the industrial electricity cost from September 2023 in northern Italy, for this condition energy costs resulted to be €21.60/year and €723.60/year for the PTC and UV respectively. Therefore, PTC energy costs are shown to be a fraction of only 3%

Table 2 Parts description for PTC and UV system (up) and water treatment conditions assumed for the economic assessment (down)

| Parabolic trough concentrator—parts description | UV treatment system—parts description |
|--|---|
| Solar 2-axis automatic tracker | UV lamp source (1000 W effective light) |
| 1×1 m special aluminum reflective mirror | Steel fixing rack |
| Steel profiles and tubes | No tracking system required—UV lamp follows quartz length |
| Polyamide machined caps and gaskets | Polyamide machined caps and gaskets |
| Quartz tube receiver (55 mm diameter × 1.2 m length) | Quartz tube receiver (55 mm diameter \times 1.2 m length) |
| Peristaltic pump | Peristaltic pump |
| Stock tank (5 L) | Stock tank (5 L) |
| Polyamide tubing (3 m) | Polyamide tubing (3 m) |
| Water treatment conditions | |
| Total volume per single treatment (L) | 5 |
| Irradiated volume (L) | 2.2 |
| Light source for 1000 W | Solar or UV |
| Initial concentration of dye (ppm) | 60 |
| Discoloration yield in 3 h (%) | 80 |
| Proposed treatment runs per year | 150 |
| Quantity of catalyst per treatment (g/L) | 1 |
| Concentration of H ₂ O ₂ per treatment (M) | 0.1 |
| Total volume treated by year (L) | 750 |

| Table 3 | Detailed yearly | investment costs | for PTC (up |) and an UV | light (down) | apparatus |
|---------|-----------------|------------------|-------------|-------------|--------------|-----------|
|---------|-----------------|------------------|-------------|-------------|--------------|-----------|

| | - | | | |
|-------------------------------------|----------|------------------|----------------------|----------------|
| Parts listing—solar PTC | Cost (€) | Lifetime (years) | Yearly cost (€/year) | Cost ratio (%) |
| Solar 2-axis automatic tracker | 8500 | 15 | 566.67 | 46.1 |
| Special aluminium reflective mirror | 300 | 2 | 150.00 | 12.2 |
| Steel profiles and tubes | 300 | 15 | 20.00 | 1.6 |
| Polyamide machined caps and gaskets | 100 | 2 | 50.00 | 4.1 |
| Quartz tube receiver | 750 | 5 | 150.00 | 12.2 |
| Peristaltic pump | 600 | 5 | 120.00 | 9.8 |
| Stock tank | 20 | 1 | 20.00 | 1.6 |
| Polyamide tubing (3 m) | 20 | 1 | 20.00 | 1.6 |
| Mechanical assembly services | 2000 | 15 | 133.33 | 10.8 |
| Total | 12,590 | | 1230.00 | 100 |
| Parts listing—UV system | Cost (€) | Lifetime (years) | Yearly cost (€/year) | Cost ratio (%) |
| UV lamp source (1000 W) | 7000 | 2 | 3500.00 | 88.4 |
| Steel fixing rack | 500 | 15 | 33.33 | 0.8 |
| Polyamide machined caps and gaskets | 100 | 2 | 50.00 | 1.3 |
| Quartz tube receiver | 750 | 5 | 150.00 | 3.8 |
| Peristaltic pump | 600 | 5 | 120.00 | 3.0 |
| Stock tank | 20 | 1 | 20.00 | 0.5 |
| Polyamide tubing (3 m) | 20 | 1 | 20.00 | 0.5 |
| Mechanical assembly services | 1000 | 15 | 66.67 | 1.7 |
| Total | 9990 | | 3960.00 | 100 |

Table 4Electricity expenses forthe PTC (up) and the UV lightapparatus (down)

| Use Equipment (h/ exp) | Power (KW) | Con- sumption (KWh) |
|---------------------------|---|--|
| | | |
| 3 | 0.1 | 0.3 |
| 3 | 0.02 | 0.06 |
| | | 0.36 |
| | | 150 |
| | | 54 |
| | | 0.4 |
| | | 21.6 |
| | | |
| 3 | 4 | 12 |
| 3 | 0.02 | 0.06 |
| | | 12.06 |
| | | 150 |
| | | 1809 |
| | | 0.4 |
| | | 723.6 |
| | Use Equipment (h/ exp) 3 3 3 3 | Use Equipment (h/ Power (KW) exp) 3 0.1 3 0.02 3 4 3 0.02 |

in comparison to the UV system, it is important to point out that the PTC advantage is an estimate based on optimal weather conditions and day-light operation, while UV systems can act as backups and substitutes for PTC systems in bad weather conditions and overnight, thus they are complementary. H_2O_2 as a sacrificial radical source and a comparison with P25 TiO₂ commercial catalyst costs are presented in Table 5. For the case of 5 L water treatment 40 ml of 30% v/v H_2O_2 is added. In the case of the photocatalyst, a loading of 1 g/L was considered as a comparison due to previously published literature that achieved the same dye equivalent

| water treatment | | | | | | |
|------------------------------------|---------------------------------------|----------------------|--|--|--|--|
| Photocatalyst | H ₂ O ₂ (0.1 M) | P25 TiO ₂ | | | | |
| Concentration (mL/L) (g/L) | 8 | 1.00 | | | | |
| Volume of experiment (L) | 5 | 5 | | | | |
| Total amount / experiment (mL) (g) | 40 | 5 | | | | |
| Price (€/mL) (€/g) | 0.02 | 0.5 | | | | |

Cost/experiment

Treatments/year

Yearly cost reagents/exp (€/year)

0.8

150

120

2.5

150

375

Table 5 Comparison costs between H₂O₂ and P25 TiO₂ for the RBB

degradations of 80% referenced by TOC analysis (Mattle and Thampi 2013). This analysis does not consider the possible reuse of the TiO₂ powder photocatalyst, as it would involve extra processing to recover the particles from the treated solution and this is not a standardized practice yet. However, reuse would in principle reduce the cost per treatment in the long term. The cost for the H_2O_2 is considered as the average price from three suppliers in tanks of 10 L (sufficient for nearly 18 months) and the P25 is based on updated values for the year 2023 from the original Evonik company. Yearly, the total cost for the acquisition of H₂O₂ is €120, and for the catalyst is €375. Notably, for a non-reuse case, the cost of using a nanoparticle photocatalyst is 56% superior to using H_2O_2 .

In Table 6, the investment costs, the electricity, and catalyst costs are put together for the water treatment along the 12-month period to represent the total annual cost. Data shows that PTC with H_2O_2 cost is $\in 1371.60$ /year, the same PTC with catalyst €1626.60/year. For the UV equipment, they are \notin 4803.60/year with H₂O₂ and \notin 5058.60/year with catalyst.

Lastly, Table 7 shows the cost per batch treatment of 5 L and an estimate per liter of water treated, PTC with H_2O_2 cost results in $\in 1.83/L$, the same PTC with catalyst $\notin 2.17/L$. For the UV equipment is $\notin 6.40/L$ with H₂O₂ and €6.74/L with catalyst.

This assessment reveals that the stronger contributions to cost composition are related to the investment costs, being from 75.6 to 89.7% in the case of PTC, and from 78.3 to 82.4% in the case of a UV system, both considering the use either of H_2O_2 or TiO₂. In specific, the constant feeding of the water treatment by H_2O_2 or adding in every step TiO₂ photocatalyst plays a minor role in the composition of treatment cost per liter.

Conclusions

This investigation focuses on the design and implementation of a solar parabolic trough concentrator, including geometric parameters and optical characteristics, as well as on its efficiency in degrading wastewater containing specific industrial dyes and emerging pharmaceutical contaminants. In the case of Remazol Brilliant Blue dye, the solar PTC application has demonstrated a degradation yield of 61% of which a remarkable 72% is due to mineralization. For the Ciprofloxacin antibiotic case, the degradation resulted in 88%, with only 8% mineralization but likely inactivation. A follow-up work with microbiology investigation is currently underway in our labs. The versatility of this method also holds promising for other pollutants, as evidenced by our ongoing studies involving two different dyes, namely Remazol Yellow and Remazol Red. Finally, our economic evaluation reveals that the solar PTC can significantly reduce treatment costs when compared to UV systems, thereby creating more favourable industrial prospects for mitigating water-related pollution issues.

Table 6 Total 12-month costs for the solar PTC (up) and UV (down) systems comprising electricity and catalysts usage

| Solar PTC | H ₂ O ₂ (0.1 M) | Ratio | P25 TiO ₂ | Ratio |
|--------------------------------|---------------------------------------|-------|----------------------|-------|
| Cost by year | €/year | % | €/year | % |
| Yearly costs of facility (YCF) | 1230.00 | 89.7 | 1230.00 | 75.6 |
| Electricity costs | 21.60 | 1.6 | 21.60 | 1.3 |
| Catalyst costs | 120 | 8.7 | 375 | 23.1 |
| Total annual costs (TAC) | 1371.60 | 100 | 1626.60 | 100 |
| UV illumination | H ₂ O ₂ (0.1 M) | Ratio | P25 TiO ₂ | Ratio |
| Cost by year | €/year | % | €/year | % |
| Yearly costs of facility (YCF) | 3960 | 82.4 | 3960 | 78.3 |
| Electricity costs | 723.60 | 15.1 | 723.60 | 14.3 |
| Catalyst costs | 120 | 2.5 | 375 | 7.4 |
| Total annual costs (TAC) | 4803.60 | 100 | 5058.60 | 100 |

| Table 7 | Cost per | liter of | treater | water v | with | PTC | (up) | and | UV | (down) |
|---------|----------|----------|---------|---------|------|-----|------|-----|----|--------|
|---------|----------|----------|---------|---------|------|-----|------|-----|----|--------|

| Solar PTC | | H_2O_2 (0.1 M) | Ratio | P25 TiO ₂ | Ratio |
|------------------------------|-----|---------------------------------------|-------|----------------------|-------|
| Costs per treatment | | €/run | % | €/run | % |
| Treatment runs | 150 | | | | |
| Cost of installation | | 8.20 | 89.7 | 8.20 | 75.6 |
| Electricity cost | | 0.14 | 1.6 | 0.14 | 1.3 |
| Catalyst cost | | 0.80 | 8.7 | 2.50 | 23.1 |
| Unit cost per treatment | | 9.14 | 100 | 10.84 | 100 |
| Liters per treatment | 5 | | | | |
| Cost per treated liter (€/L) | | 1.83 | | 2.17 | |
| UV illumination | | H ₂ O ₂ (0.1 M) | Ratio | P25 TiO ₂ | Ratio |
| Costs per treatment | | €/run | % | €/run | % |
| Treatment runs | 150 | | | | |
| Cost of installation | | 26.40 | 82.4 | 26.40 | 78.3 |
| Electricity cost | | 4.82 | 15.1 | 4.82 | 14.3 |
| Catalyst cost | | 0.80 | 2.5 | 2.50 | 7.4 |
| Unit cost per treatment | | 32.02 | 100 | 33.72 | 100 |
| Liters per treatment | 5 | | | | |
| Cost per treated liter (€/L) | | 6.40 | | 6.74 | |

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Availability of data and materials Data can be made available on request.

Declarations

Conflict of interest The authors declare that they have no competing interests.

Ethics approval and consent to participate Not applicable.

Consent for publication Not applicable.

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